

## Supplementary material

### **Sludge-derived novel Fe<sub>3</sub>O<sub>4</sub>-MgO@C composites for tetracycline abatement from wastewater using sustainable bio-electro-Fenton and nutrients recovery with co-precipitation: A waste-to-resource strategy**

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**Table S1.** Characteristics of raw municipal wastewater

### **Text S1. Characterization of catalyst**

Electrochemical analyses, including electrochemical impedance spectroscopy (EIS), linear sweep voltammetry (LSV), cyclic voltammetry (CV), and Tafel analyses, were conducted using an electrochemical workstation (Metrohm, Switzerland). The workstation was comprised of a three-electrode system with glassy carbon, platinum, and Ag/AgCl serving as working, counter, and reference electrodes, respectively. All electrochemical measurements were carried out in a 50 mM Na<sub>2</sub>SO<sub>4</sub> solution at a scan rate of 0.01 mV/s. Morphological analysis of the catalyst was performed using scanning electron microscopy (SEM) equipment (Carl Zeiss, USA). Fourier-transform infrared spectroscopy (FTIR) was conducted within the wavenumber range of 400 to 4000 cm<sup>-1</sup> using Thermo Fisher Scientific Instruments, USA equipment. X-ray diffraction (XRD) was utilized to investigate the structural characteristics and composition of the waste-derived catalyst and sludge. The XRD analysis involved scanning 2θ from 10 to 80° at a scanning rate of 2° min<sup>-1</sup> and a step size of 0.02 using a Bruker D2 Phaser instrument from Germany.

The surface area of the composite catalyst was measured using a multipoint Brunauer–Emmett–Teller (BET) analyser from BEL, Japan, Inc. X-ray photoelectron spectroscopy (XPS) was employed to determine the structural characteristics of the catalyst, utilizing equipment from ULVAC Physical Electronics, USA. Additionally, the magnetic properties of the waste-derived catalyst were analysed using a vibrating sample magnetometer (VSM) from Lake Shore Cryotronics, USA.

### **Text S2. Economic analysis**

The economics of electro-Fenton operation was aimed at assessing the real-scale capability of the process. It is established on the operating cost ( $OC_{\text{total}}$ ) needed to eliminate contaminants by unit order of magnitude per m<sup>3</sup> of contaminated water, which is also recognized as  $OC_{\text{total}}$  per order. The  $OC_{\text{total}}$  includes the catalyst cost, chemical cost, and electricity cost, which were determined by calculating the energy consumption ( $W$ ) and electrical energy per order ( $EEO$ ) as per Eq. (S1) – Eq. (S5)

$$W \left( \frac{kWh}{m^3} \right) = \frac{UIT}{V} \quad (S1)$$

$$EEO \left( \frac{kWh}{m^3 \log} \right) = \frac{W}{\log \left( \frac{C_o}{C} \right)} \quad (S2)$$

$$OC_{total} = \text{Catalyst cost} + \text{Chemical cost} + \text{Electricity cost} \quad (S3)$$

Where,  $U$  is the voltage (V),  $I$  is the imposed current (A),  $T$  is the treatment duration (h),  $V$  is the volume ( $m^3$ ), and  $C_o$  and  $C$  are the initial and final concentrations of contaminants (mg/L), respectively.

$$\text{Chemical cost } (\$/m^3) = \text{Electrolyte utilized } (kg/m^3) \times \text{cost of electrolyte } (\$/kg) \quad (S4)$$

$$\text{Electricity cost } (\$/m^3) = \text{EEO } (kWh/m^3/\log) \times \text{per unit cost of electrical energy } (\$/kWh) \quad (S5)$$

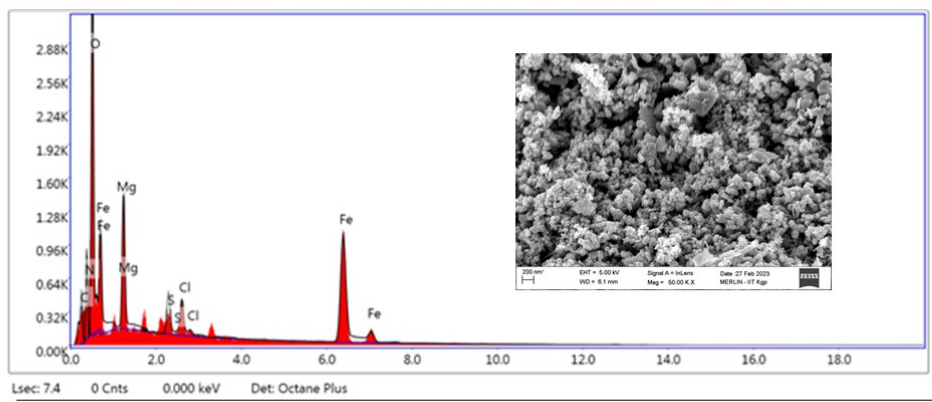
Additionally, the mineralization current efficiency (MCE%) was computed with  $\Delta TOC$  (mg/L) using Eq. (S6)

$$MCE (\%) = \frac{nFV\Delta TOC}{Cmt} \quad (S6)$$

Where,  $n$  is the total number of electrons consumed by SDS for mineralization,  $t$  is the treatment duration (h),  $V$  is the wastewater volume (L),  $F$  indicates the Faraday's constant ( $96485 \text{ C mol}^{-1}$ ),  $c$  represents the conversion factor ( $4.32 \times 10^7$ ), and  $I$  is the external applied current (A).

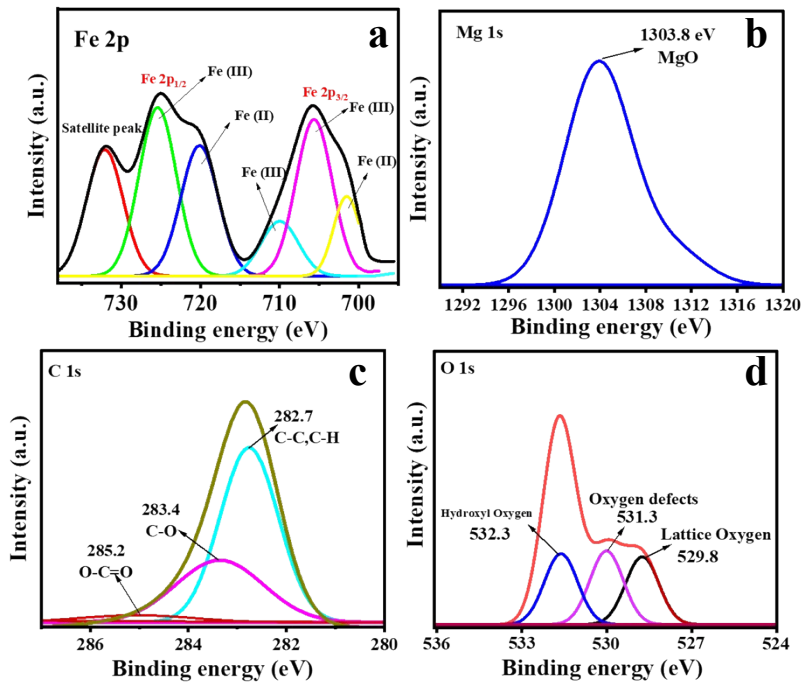
### Text S3: Synthetic urine

The synthetic urine contained 1.36 g/L  $\text{NaCH}_3\text{COO} \cdot 3\text{H}_2\text{O}$ , 0.74 g/L  $\text{KCl}$ , 0.58 g/L  $\text{NaCl}$ , 0.68 g/L  $\text{KH}_2\text{PO}_4$ , 0.87 g/L  $\text{K}_2\text{HPO}_4$ , 0.28 g/L  $\text{NH}_4\text{Cl}$ , 0.1 g/L  $\text{MgSO}_4 \cdot 7\text{H}_2\text{O}$ , 0.1 g/L  $\text{CaCl}_2 \cdot 2\text{H}_2\text{O}$  and 0.1 mL/L of a trace element mixture.

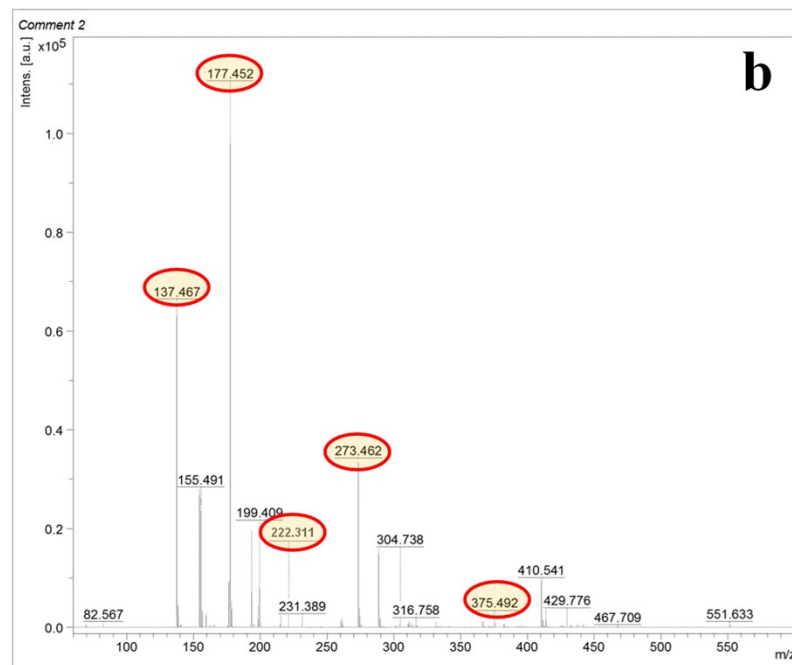
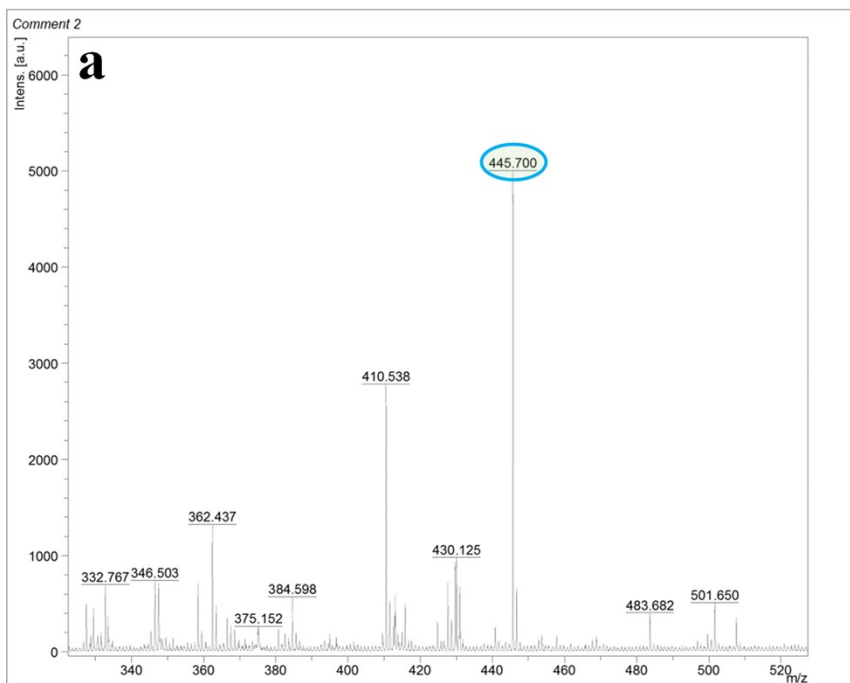


Element	Weight %	Atomic %	Net Int.	Error %	Kratio
C K	4.07	9.68	364.36	7.67	0.0199
N K	7.44	15.15	652.18	6.03	0.0419
O K	18.19	32.46	3361.72	4.33	0.1098
MgK	7.21	8.46	1599.65	5.36	0.0389
S K	2.84	2.53	537.76	5.57	0.0191
ClK	3.17	2.55	524.20	7.69	0.0207
FeK	57.08	29.17	2390.25	3.76	0.3868

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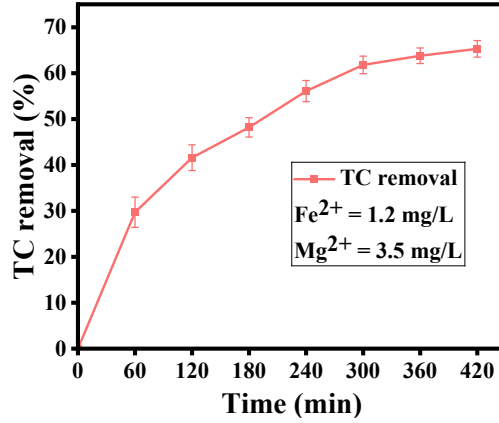


Fig. S4. Degradation of TC antibiotic in homogeneous BEF process (TC = 10 mg/L and pH = 7.0)

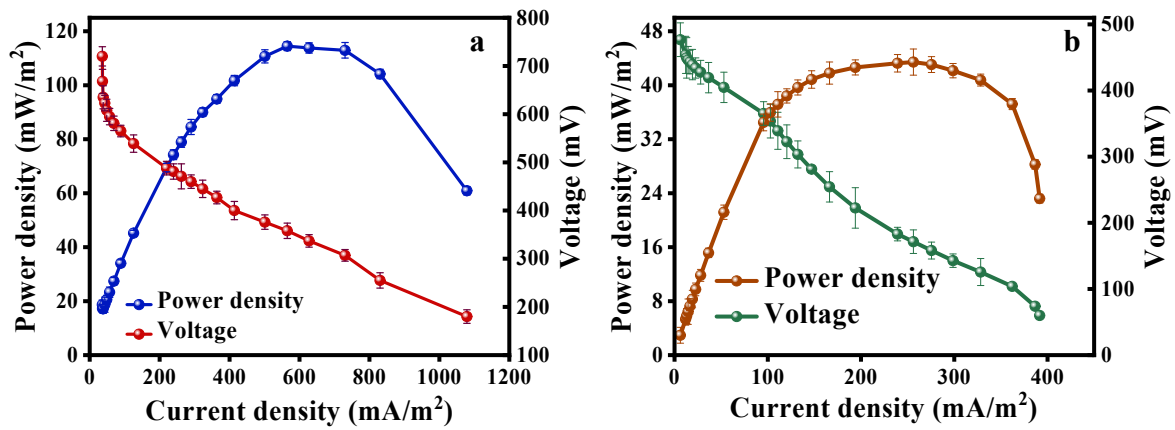
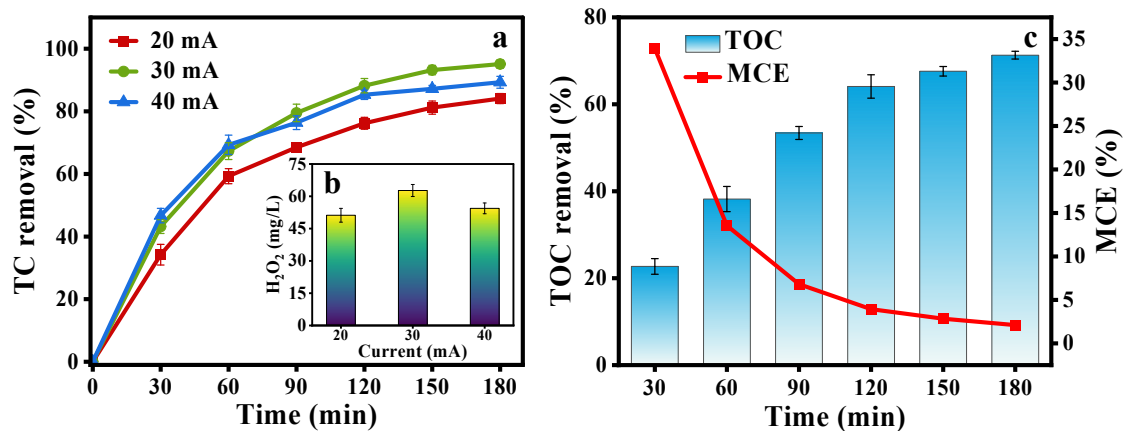
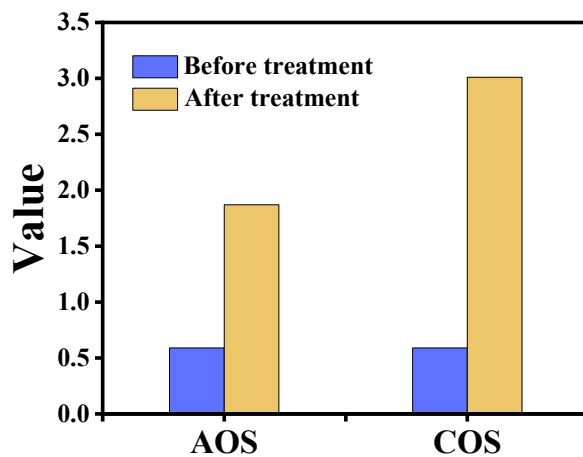


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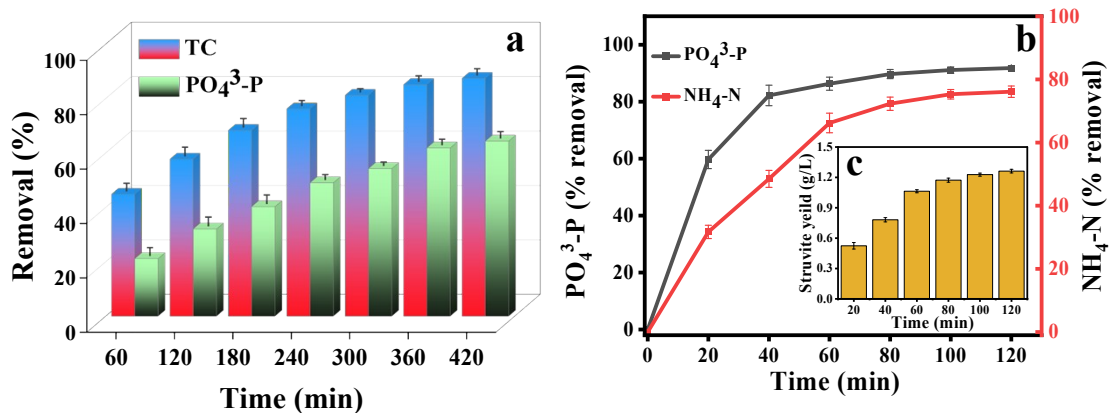




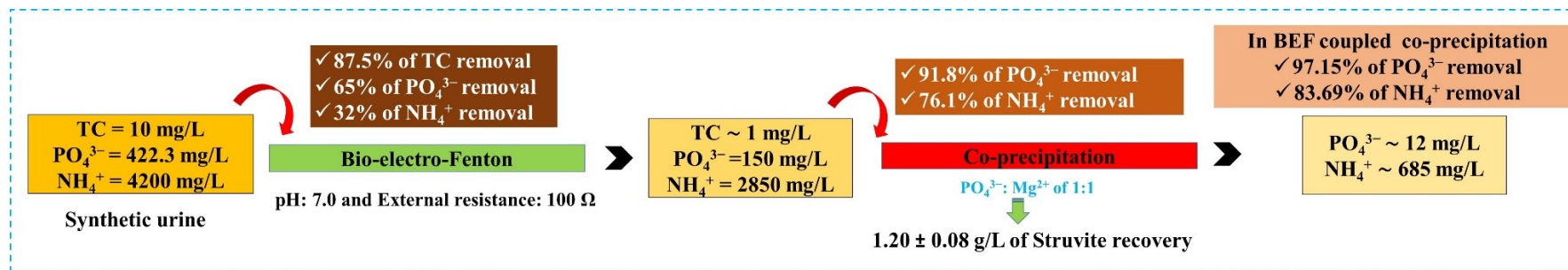
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**Fig. S9.** Schematic of BEF and co-precipitation revealing contaminant removal in each process

**Table S1.** Characteristics of raw municipal wastewater

<b>Parameters</b>	<b>Value</b>
COD (mg L <sup>-1</sup> )	295 ± 15
pH	7.1 ± 0.2
TOC (mg L <sup>-1</sup> )	130 ± 10
Turbidity (NTU)	97.5 ± 6.0
Cl <sup>-</sup> (mg L <sup>-1</sup> )	70 ± 10
Conductivity (μS cm <sup>-1</sup> )	520 ± 20
SO <sub>4</sub> <sup>2-</sup> (mg L <sup>-1</sup> )	15.5 ± 3.0
SDS (mg L <sup>-1</sup> )	3.3 ± 0.1