1	Supporting information
2	Regulated Dual Defects of Ligand Defects and Lattice Defects
3	in UIO-66 for Ultra trace Simultaneous Detection and Removal
4	of Heavy Metal Ions
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54 1. Experimental

55 1.1 Materials

All chemicals were analytically pure and were used as purchased, without any further purifications. Zirconyl Chloride Octahydrate, 98% was purchased from Shanghai Adamas Reagent Co., Ltd. Absolute ethanol, Cerium nitrate hexahydrate 99.5%, Trifluoroacetic acid AR, p-Phthalic acid 99% were purchased from Shanghai Macklin Biochemical Technology Co., Ltd. Deionized water was used throughout the electrochemical measurements and materials preparation.

62 1.2 Material characterization

The synthesized D-D-UIO-66 materials were characterized by the powder X-ray 63 diffraction (XRD; Rigaku XRD-6100, Japan); scanning electron microscopy (SEM, 64 Hitachi SU8010, Japan) with energy dispersive X-ray spectroscopy (EDX) and 65 transmission electron microscopy (TEM; FEI Talos F200X G2, USA); X-ray 66 photoelectron spectroscopy (XPS; Thermo Scientific K-Alpha, USA) was conducted 67 using monochromatic Al Ka X-ray source; The specific surface areas were obtained 68 from N₂ adsorption desorption isotherms at 77 K with BET surface area analyzer (BET; 69 Micromeritics, ASAP 2460, GA, USA); Thermogravimetric analysis (TG) 70 measurements were made with a DSC/DTA-TG unit (Rigaku TG/DTA8122, Japan) 71 under N_2 atmosphere from 30 °C to 800 °C at heating rate of 10 °C min^-1; The 72 concentration of HMIs was determined by a ContrA 700 (Analytik Jena, Germany) 73 Inductively Coupled Plasma-Atomic Emission Spectrometry (ICP-AES). 74

75 1.3Inkjet printing

The ink was filtrated through a filter with a pore size of 0.45 μm before printing.
Picolitre drops of the ink were ejected from a dope-on-demand inkjet system RDPMF100 (MicroFab Technologies) onto the glassy carbon electrode (GCE). Printing
was performed at a voltage pulse of 27 V with a droplet velocity of 1.24 m s⁻¹

80 1.4 HMIs detection

All electrochemical experiments were performed on an electrochemical 81 workstation (CHI 760E, CH Instruments Inc., Shanghai, China) with a conventional 82 three-electrode system, where the bare or D-D-UIO-66/GCE modified electrode, 83 Ag/AgCl and platinum wire electrode served as the working, reference, and auxiliary 84 electrodes, respectively. The differential pulse anodic stripping voltammetry (DPASV) 85 measurement was utilized for the observation of electrochemical detection behavior 86 toward Pb(II), Cd(II), Cu(II), and Hg(II) under optimized experimental conditions. The 87 actual water is first filtered to remove insoluble impurities, then diluted with a 0.1 M 88 HAc-NaAc buffer solution to form a real water sample with a volume ratio of 1:9. 89

90 1.5 HMIs adsorption

Adsorption isotherms. During an isotherm adsorption experiments, 20 mg adsorbents were dispersed in 20 mL Pb(II) aqueous solution with initial concentrations ranged from 100 to 1000 mg J L⁻¹. The adsorption process performed in a shaker at 200 rpm for 700 minutes at 25 °C. The adsorption capacity of Pb(II) on four adsorbents was calculated using Eq. (1):

$$Q_e = \frac{V(C_0 - C_e)}{m}$$

(1)

Where Q_e is the equilibrium adsorption capacity (mg L⁻¹), C_0 and C_e are the initial and equilibrium concentration of Pb(II) (mg L⁻¹), respectively, *m* is the mass of adsorbents (g), *V* is the volume of Pb(II) solution (L). The isotherm adsorption dates were analyzed using Langmuir Eq. (2)
and the Freundlich Eq. (3) isotherm model at 25 °C

$$Q_e = \frac{K_L Q_m C_e}{I + K_L C_e} \quad (2)$$

100

101
$$Q_e = K_F C_e^{l/n}$$
 (3)

Where C_e (mg g⁻¹) is the equilibrium concentration of the Pb(II) solution, Q_e (mg g⁻¹) is the equilibrium adsorption capacity of the adsorbent, Q_m (mg g⁻¹) is the theoretical maximum adsorption capacity, K_L (L mg⁻¹) represents Langmuir constant, K_F (mg^{-1/n} L^{-1/n} g⁻¹) and *n* represents Freundlich constant.

Adsorption kinetics. Adsorption kinetics was conducted by adding 100 mg of the four adsorbents into a 100 mL Pb(II) solution with initial concentrations 1000 mg L⁻¹ at 25 °C. The kinetic dates were examined by the pseudo-first-order Eq. (4) and pseudo-second-order model Eq. (5):

110
$$ln(Q_e - Q_t) = lnQ_m - k_l t$$
 (4)

$$\frac{t}{Q_t} = \frac{l}{k_2 Q_m^2} + \frac{t}{Q_m}$$
(5)

Where Q_t is the adsorption capacity of the adsorbent at time t and Q_m (mg g⁻¹) and at equilibrium, respectively, k_1 (min⁻¹) and k_2 (g mg⁻¹ min⁻¹) are the corresponding rate constants.

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117 2. Results and discussion





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Figure S1 HRTEM and Fast Fourier Transform of D-D-UIO-66



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Figure S2 Full XPS spectra of UIO-66, D-UIO-66 and D-D-UIO-66







127 Figure S3 Simultaneous detection of D-UIO-66 for Cd(II), Pb(II) Cu(II) and Hg(II).







 $[Fe(CN)_6]^{3-/4-}$ and 0.1 M KCl and the linear relationship between the anodic peak

currents and the square root of scan.



Figure S6 CV curves of D-UIO-66 /GCE in the solution containing 5.0 mM
[Fe(CN)₆]^{3-/4-} and 0.1 M KCl and the linear relationship between the anodic peak
currents and the square root of scan.











Figure S7 The reproducibility of D-D-UIO-66.









151

3

0

1

3 0.15

2

0.20

Figure S8 The stability of D-D-UIO-66.

0.25

3

Hg(II)

0.30

Potential (V)

4

Repeat Days

0.35

5

0.40

0.45

6







Figure S9 the DPASV curves in Shuxiang Lake in Shihezi, China for the Cd(II),

157 Pb(II) Cu(II) and Hg(II).





(Tongling, China) for the Cd(II), Pb(II) Cu(II) and Hg(II)





Figure S11 The comparison of Cd(II), Pb(II) Cu(II) and Hg(II) sensitivity in

different types of water





Figure S12 O 1s XPS Spectra of D-D-UIO-66 before and after adsorption.



		Sensitivity	Detection limit	References
Electrodes	Method	$(\mu A \; \mu M^{-1})$	(nM)	
Bi/Bi2O3@C	DPASV	3.35	6.3	1
BCN-Nafion/GCE	SWAS V	0.509	0.9	2
Fe3O4@MPC-2/GCE	DPASV	22.3	12.1	3
Fe@YAU-101/GCE	DPASV	0.596	33.3	4
e-CuFe-PBA/GCE	DPASV	24.915	28.7	5
α-ΜοΟ	SWAS V	0.033	9.72	6
g-C3N4-P(Ani-Py)-PAAM	DPASV	1.2906	14.84	7
Cu-Co3O4 MCNS	DPASV	20.59	20.9	8
ALA/pDA/rGO	DPASV	15.41	13.774	9
D-D-UIO-66 (This work)	DPASV	20.204	5.965	

Table S1 Compared with other materials Pb(II) detection performance

Table S2 Comparison of individual and simultaneous detection

	HMIs	Sensitivity	\mathbb{R}^2	•
		(μA μM ⁻¹)		
Simultaneous	Cd (II)	10.092	0.959	•
detection	Pb (II)	15.209	0.996	
	Hg (II)	1.347	0.916	
	Cu (II)	2.829	0.980	
Individual	Cd (II)	12.838	0.993	
detection	Pb (II)	20.204	0.989	
	Hg (II)	1.668	0.969	

Cu (II)	2.387	0.983

177 Table S3. Langmuir and Freundlich adsorption parameters for Pb(II) on the four adsorbents at 298

178				К.		
	Lang	gmui	ir isotherm pa	rameters	Freundlich isother	m parameters
Adsorbe	nts Q_n	n	K_L (*10 ⁻³)	R ²	K_F	R ²
	(mg	g-1)	(L mg ⁻¹)	ĸ	$(mg^{1-(1/n)}L^{1/n}g)$	K
D-UIO-	66 416.	85	0.0001	0.9864	0.301	0.9440
D-D-UIO	-66 667.	04	0.0156	0.9968	0.473	0.9848
170						

179

181

180 Table S4. Estimates of parameter values for the adsorption of Pb(II) on adsorbents at 298 K

according to the pseudo-first-order and pseudo-second-order models.

	pseudo-first-order			pseudo-second-order		
Adsorbents	Q_e	K_{l}	D 2	Q_m	K_2 (*10 ⁻⁴)	D 2
	(mg g ⁻¹)	(min ⁻¹)	K ²	(mg g ⁻¹)	(g mg ⁻¹ min ⁻¹)	κ-
D-UIO-66	349.94	0.126	0.982	364.823	6.798	0.991
D-D-UIO-66	385.92	0.128	0.969	402.40	6.184	0.973

182

183 Table S5 Comparison of LOD and theoretical adsorption capacity for Pb(II) with other

1	84	
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bifunctional materials

 Absorbent	Q _m (mg g ⁻	Detection limit	References	-
	1)	(nM)		
 BUC-77	425	33.36	10	
IIMB	124.07	4.58	11	
FSH-6	265.9	0.289	12	

NBW	211.6	0.627	13
CDs	183	-	14
D-D-UIO-66	667.04	5.965	This work

186 Table S6 Comparison of theoretical adsorption capacity for Pb(II) with other monofunctional

187		materials	
	Absorbent	$Q_m \left(mg \; g^{-1}\right)$	References
	SALDETA@CPTMS@Fe ₃ O	415.5	15
	4		
	Fe-LAA	508.2	16
	A/M-CDMOF	414.2	17
	MOF-DFSA	349.09	18
	D-D-UIO-66	667.04	This work

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